

Altitude ammonia-oxidizing bacteria and archaea in soils of Mount Everest

Li-Mei Zhang¹, Mu Wang², James I. Prosser³, Yuan-Ming Zheng¹ & Ji-Zheng He¹

¹State Key Laboratory of Urban and Regional Ecology, Research Centre for Eco-environmental Sciences, Chinese Academy of Sciences, Beijing, China; ²Agricultural and Animal Husbandry College of Tibet, Tibet, China; and ³Institute of Biological and Environmental Sciences, University of Aberdeen, Aberdeen, UK

Correspondence: Ji-Zheng He, State Key Laboratory of Urban and Regional Ecology, Research Centre for Eco-environmental Sciences, Chinese Academy of Sciences, Beijing 100085, China. Tel.: +86 10 6284 9788; fax: +86 10 6292 3563; e-mail: izhe@rcees.ac.cn

Received 6 April 2009; revised 23 July 2009; accepted 21 August 2009. Final version published online 23 September 2009

DOI:10.1111/j.1574-6941.2009.00775.x

Editor: Christoph Tebbe

Keywords

alpine soil; ammonia oxidizer; *amo*A gene; community shift; relative abundance; Tibetan Plateau.

Abstract

To determine the abundance and distribution of bacterial and archaeal ammonia oxidizers in alpine and permafrost soils, 12 soils at altitudes of 4000-6550 m above sea level (ma.s.l.) were collected from the northern slope of the Mount Everest (Tibetan Plateau), where the permanent snow line is at 5800–6000 m a.s.l. Communities were characterized by real-time PCR and clone sequencing by targeting on amoA genes, which putatively encode ammonia monooxygenase subunit A. Archaeal amoA abundance was greater than bacterial amoA abundance in lower altitude soils (< 5400 m a.s.l.), but this situation was reversed in higher altitude soils (\geq 5700 m a.s.l.). Both archaeal and bacterial *amoA* abundance decreased abruptly in higher altitude soils. Communities shifted from a Nitrosospira amoA cluster 3a-dominated ammonia-oxidizing bacteria community in lower altitude soils to communities dominated by a newly designated Nitrosospira ME and cluster 2-related groups and Nitrosomonas cluster 6 in higher altitude soils. All archaeal amoA sequences fell within soil and sediment clusters, and the proportions of the major archaeal amoA clusters changed between the lower altitude and the higher altitude soils. These findings imply that the shift in the relative abundance and community structure of archaeal and bacterial ammonia oxidizers may result from selection of organisms adapted to altitude-dependent environmental factors in elevated soils.

Until recently, chemolithoautotrophic ammonia-oxidizing bacteria (AOB) within the *Proteobacteria* were thought to be the major microorganisms performing ammonia oxidation, the first and rate-limiting step of nitrification. Metagenomic studies first suggested that ammonia oxidation was potentially driven by members within the domain archaea that contained putative ammonia monooxygenase subunits (*amoA*, *amoB* and *amoC*) (Venter *et al.*, 2004; Treusch *et al.*, 2005). The first successful isolation of a mesophilic crenarchaeon, *Nitrosopumilus maritimus*, growing autotrophically with ammonia oxidation (Könneke *et al.*, 2005). Given the functional significance and conserved phylogeny, the *amoA* gene was used as a molecular marker for both

AOB and these putative ammonia-oxidizing archaea (AOA). Subsequent studies have demonstrated that archaeal *amoA* genes are ubiquitous in marine environments (Francis *et al.*, 2005; Wuchter *et al.*, 2006), wastewater bioreactors (Park *et al.*, 2006), terrestrial hot springs (Weidler *et al.*, 2007; Reigstad *et al.*, 2008) and soils (He *et al.*, 2007; Nicol *et al.*, 2008; Shen *et al.*, 2008). Active expression of archaeal *amoA* gene has also been detected in natural environments (Lam *et al.*, 2007; Tourna *et al.*, 2008).

Archaeal *amo*A genes are more abundant than those of bacteria in various environments. In marine environments, archaeal *amo*A gene abundance exceeded that of bacteria by up to four orders of magnitude (Wuchter *et al.*, 2006; Lam *et al.*, 2007; Mincer *et al.*, 2007). In a study of 12 pristine and agricultural soils crossing three climate zones, archaeal *amo*A gene was up to 3000 times that of bacterial *amo*A

and correlated with Crenarchaeota-specific lipids (Leininger et al., 2006). Consistently higher abundance of archaeal amoA genes has also been reported in Chinese acid and alkaline soils with different fertilization regimes (He et al., 2007; Shen et al., 2008) and in Scottish agricultural plots (Nicol et al., 2008). The numerical dominance of archaeal over bacterial amoA genes in natural environments suggested its putative greater ecological function in global nitrogen cycling, but the presence or the high abundance of a functional gene does not mean that the function is operating (Prosser & Nicol, 2008). Moreover, some recent studies reported greater abundance of AOB amoA genes in estuarine sites (Caffrey et al., 2007), in coastal sediments with increasing salinity (Mosier & Francis, 2008; Santoro et al., 2008) and in anoxic sediments (Jiang et al., 2009). It is important to understand whether there are more habitats in which AOB dominate, especially in soils, and the physical and chemical factors that determine the relative abundance and diversity of these two distinct groups of ammonia oxidizers.

A number of studies have shown that the distribution, community composition and abundance of AOB and AOA are influenced considerably by temperature (Avrahami *et al.*, 2003), pH (Nicol *et al.*, 2008), simulated global changes (Horz *et al.*, 2004), salinity (Mosier & Francis, 2008; Santoro *et al.*, 2008) and fertilization regime (He *et al.*, 2007; Shen *et al.*, 2008). Distinct changes in the AOA community composition have also been observed in a soil profile, in which AOB were not detected (Hansel *et al.*, 2008), and a marked decrease in the abundance of archaeal *amoA* gene was found with increasing depth, from subsurface waters to 4000 m depth in the North Atlantic Ocean (Agogue *et al.*, 2008). However, AOB and AOA communities in high-altitude soils have not been investigated.

The Tibetan Plateau is the Earth's largest $(2 \times 10^6 \text{ km}^2)$ and highest [mean altitude 4500 m above sea level (m a.s.l.)] plateau. Tectonic uplifting of the Tibetan Plateau has led to a unique system, with snow cover, high UV exposure and lower oxygen and nutrient concentrations. As a result, microbial communities in topsoil layers encounter extreme conditions that may lead to unique survival adaptations and differences in the community composition. The Tibetan Plateau is also one of the most special regions sensitive to global climate change (Wu et al., 2007) and the global warming and elevated CO₂ concentration may have a complex impact on below-ground microorganisms (Xu et al., 2009). In the present study, real-time PCR and clone-library sequencing were used to characterize the distribution and community composition of AOA and AOB in these alpine soils, where the altitudedependent environmental factors and changing global climate may select for particular populations of ammonia oxidizers.

Materials and methods

Sampling sites and soil collection

During the fourth expedition on Mount Everest of the Chinese Academy of Sciences from April to June 2005, 12 soil samples (M1–M12) were collected from the northern slope of Mount Everest and adjoining areas (Table 1). M1–M3 are field or farmland soils in the Rikaze area, Tibet, at an altitude of 4000 m a.s.l. M4–M12 are bare soils with no visible plant development (Fig. 1). The permanent snow line is at 5800–6000 m a.s.l. For each site, three subsamples were removed from the soil surface (0–10 cm depth) and were mixed after removal of large stones and snow. Samples were transported to a refrigerator with a temperature of -10 °C in 2 days after collecting and stored until further analysis. Soils were passed through a 2.0-mm sieve before use.

Soil chemical analysis

Soil pH was determined with a soil to water ratio of 1:10. Soil organic matter was determined using the $K_2Cr_2O_7$ oxidation method, and total nitrogen was determined using the Kjeldahl method (Bremner, 1996).

DNA extraction

DNA was extracted from 0.8 to 1.0 g (fresh weight) of soil using the FastDNA SPIN Kit for Soil (Q BIOgene Inc.,

 Table 1. Soil characteristics for samples along the Mount Everest on the

 Tibetan Plateau

| Altitudo | | | Organic | Total N |
|----------|---|--|--|--|
| (m) | Location | рН | (g kg ⁻¹) | (g kg ⁻¹) |
| 4000 | Lazi field soil | 8.6 | 8.4 | 0.51 |
| 4000 | Rikaze farmland soil | 8.7 | 10.8 | 0.66 |
| 4000 | Dazhuka field soil | 8.6 | 12.6 | 0.49 |
| 5350 | N28 08.385 | 9.1 | 4.1 | 0.14 |
| | E86 51.533 | | | |
| 5400 | N28 08.336 | 9.0 | 6.3 | 0.21 |
| | E86 51.602 | | | |
| 5700 | N28 05.376 | 9.1 | 4.3 | 0.06 |
| | E86 54.469 | | | |
| 5850 | N28 04.882 | 9.0 | 4.2 | 0.07 |
| | E86 55.256 | | | |
| 6000 | N28 03.892 | 9.0 | 2.9 | 0.06 |
| | E86 55.605 | | | |
| 6150 | N28 03.030 | 9.1 | 3.7 | 0.09 |
| | E86 56.294 | | | |
| 6300 | N28 02.341 | 9.0 | 4.4 | 0.11 |
| | E86 56.686 | | | |
| 6450 | N28 01.773 | 9.0 | 4.4 | 0.11 |
| | E86 56.337 | | | |
| 6550 | N28 01.570 | 9.0 | 5.9 | 0.09 |
| | E86 55.987 | | | |
| | Altitude (m) 4000 4000 5350 5400 5700 5850 6000 6150 6300 6450 6550 | Altitude Location 4000 Lazi field soil 4000 Rikaze farmland soil 4000 Dazhuka field soil 4000 Dazhuka field soil 4000 Dazhuka field soil 5350 N28 08.385 5850 E86 51.533 5400 N28 08.336 5850 N28 05.376 5850 N28 04.882 5850 N28 04.882 586 55.256 6000 6000 N28 03.392 586 55.256 605 6150 N28 03.030 586 56.294 6300 6300 N28 02.341 586 56.686 6450 6450 N28 01.773 586 56.337 5550 6550 N28 01.570 586 55.987 | Altitude pH (m) Location pH 4000 Lazi field soil 8.6 4000 Rikaze farmland soil 8.7 4000 Dazhuka field soil 8.6 4000 Dazhuka field soil 8.6 5350 N28 08.385 9.1 E86 51.533 1 1 5400 N28 08.336 9.0 E86 51.602 1 1 5700 N28 05.376 9.1 E86 51.602 1 1 5700 N28 03.892 9.0 E86 55.256 1 1 6000 N28 03.030 9.1 E86 56.294 1 1 6300 N28 02.341 9.0 E86 56.686 1 1 6450 N28 01.773 9.0 E86 56.337 1 1 6550 N28 01.570 9.0 E86 55.987 1 1 | Altitude Organic matter (m) Location pH (g kg ⁻¹) 4000 Lazi field soil 8.6 8.4 4000 Rikaze farmland soil 8.7 10.8 4000 Dazhuka field soil 8.6 12.6 5350 N28 08.385 9.1 4.1 5350 N28 08.336 9.0 6.3 5400 N28 08.336 9.0 6.3 5400 N28 08.336 9.0 6.3 5700 N28 05.376 9.1 4.3 5850 N28 04.882 9.0 4.2 5850 N28 03.302 9.0 2.9 6000 N28 03.303 9.0 2.9 6150 N28 03.303 9.1 3.7 6300 N28 03.303 9.1 3.7 6300 N28 03.303 9.0 4.4 86 56.294 6450 N28 01.773 9.0 4.4 6450 N28 01.773 9.0< |



Fig. 1. Location of soils sampled in the Mount Everest, on the Tibetan Plateau.

Carlsbad, CA) with a bead beating time of 20 s and a speed setting of 5.5 m. Extracted DNA was checked on a 1% agarose gel and the concentration was determined using a Nanodrop[®] ND-1000 UV-Vis Spectrophotometer (NanoDrop Technologies, Wilmington, DE).

210

Quantification of amoA genes by real-time PCR

Primer pairs amoA1F/amoA2R (Rotthauwe et al., 1997) and Arch-amoAF/Arch-amoAR (Francis et al., 2005) were used for real-time PCR quantification of bacterial and archaeal amoA genes, respectively. Real-time PCR was performed on an iCycler iQ 5 thermocycler (Bio-Rad). The 25 µL reaction volume contained 12.5 µL SYBR® Premix Ex TaqTM (Ta-KaRa Bio Inc., Shiga, Japan), 0.4 mg mL⁻¹ bovine serum albumin, 400 nM of each AOB primer or 200 nM each of AOA primer and 2 µL of 10-fold diluted or undiluted extracted DNA (1-10 ng) as a template. Three replicates were analyzed for each sample. Amplifications were carried out as follows: 95 °C for 1 min, followed by 40 cycles of 10 s at 95 °C, 30 s at 55 °C for AOB or 53 °C for AOA, 1 min at 72 °C and plate read at 83 °C. Melting curve analysis was performed at the end of PCR runs to check the specificity of the products. PCR products amplified from extracted DNA

with the primers for real-time PCR assays were gel-purified and ligated into the pGEM-T Easy Vector (Promega, Madison), and the resulting ligation products were transformed into Escherichia coli JM109 competent cells following the manufacturer's instructions. After reamplification with the vector-specific primers T7 and SP6, the positive clones were selected to extract plasmid DNA using a MiniBEST Plasmid Purification Kit (TaKaRa Bio Inc.) and used as amoA gene standards. The plasmid DNA concentration was determined on a Nanodrop[®] ND-1000 UV-Vis Spectrophotometer (NanoDrop Technologies), and amoA gene copy number was calculated directly from the concentration of extracted plasmid DNA. Tenfold serial dilutions of a known copy number of the plasmid DNA were subjected to real-time PCR in triplicate to generate an external standard curve. PCR efficiency and correlation coefficients for standard curves were 95.4% and $r^2 = 0.996$ for AOB and 93.4% and $r^2 = 0.996$ for AOA.

Cloning and sequence analysis of bacterial and archaeal *amo*A genes

PCR products of bacterial and archaeal *amoA* genes amplified with the primers for the real-time PCR assays were purified and cloned, as described above, for each site, generating 12 clone libraries for bacterial and 12 for archaeal *amoA* genes. Clones were screened for inserts of the expected size with the vector-specific primers T7 and SP6. For restriction fragment length polymorphism (RFLP) assay, approximately 60 positive clones from each clone library were digested with restriction endonuclease MboI (Takara Bio Inc.). Digested DNA fragments were separated by electrophoresis on a 2% agarose gel and imaged using a GBOX-HR Gel Documentation System (Syngene, UK) after ethidium bromide staining.

The RFLP patterns were grouped and one to three clones representing unique band types were sequenced. Sequences were subjected to homology analysis using the software DNAMAN version 6.0.3.48 (Lynnon Biosoft). Sequences of chimeric origin were checked by partial treeing analysis and compared with GenBank sequences using BLASTN searches. For each clone library, for sequence types that exhibited > 98% identity to each other, only one representative was used for construction of the tree. The GenBank sequences most similar to clone sequences in this study and reference sequences for defining clusters were included in phylogenetic tree construction. Phylogenetic analysis based on nucleotide sequences was performed using MEGA version 4.0 (Tamura et al., 2007) and a neighbor-joining tree was constructed using Kimura two-parameter distance with 1000 replicates to produce bootstrap values (1000 replicates).

The sequences determined in this study were deposited in the GenBank database and assigned accession numbers from FJ853214 to FJ853366.

Statistical analyses

*amo*A gene abundance data were log-transformed to provide variance homogeneity. Statistical analyses were performed using spss version 13.0. Paired-samples *t*-tests were used to compare archaeal and bacterial *amo*A gene copy numbers and independent-sample *t*-tests were used to compare archaeal or bacterial *amo*A gene copy numbers at lower

altitudes and at higher altitudes. Bivariate correlations were carried out to link different parameters. P < 0.05 was considered to be significant.

Results

Soil properties

Three soils (M1–M3) at an altitude of 4000 m a.s.l. were sampled from field or farmland with pH values of 8.6–8.7. Organic matter in these three soils varied between 8.4 and 12.6 g kg⁻¹ soil and total nitrogen varied between 0.51 and 0.66 g kg⁻¹ soil (Table 1). Soils (M4–M12) at altitudes above 5000 m a.s.l. were less well developed with a high content of gravel, and pH was in the range 9.0–9.1. The organic matter content in these soils ranged from 2.9 to 6.3 g kg⁻¹ soil, and total nitrogen from 0.06 to 0.21 g kg⁻¹ soil. These values are much lower than in M1–M3 soils and showed no significantly negative correlation with altitude (Table 2).

AOA and AOB abundance

The abundance of archaeal and bacterial amoA genes decreased abruptly in soils above 5400 m a.s.l. and significant differences in archaeal and bacterial amoA abundances were observed between lower altitude (M1–M5, < 5400 m a.s.l.) and higher altitude (M6–M12, \geq 5700 m a.s.l.) soils (d.f. = 10, P < 0.05; Fig. 2a). Differences were also seen in the relative abundances of archaeal and bacterial amoA genes. Archaeal amoA abundance in lower altitude soils $(5.17 \times 10^7 - 3.79)$ $\times 10^8 \,\mathrm{g}^{-1}$ soil) was significantly higher than bacterial *amo*A abundance $(6.97 \times 10^5 - 1.59 \times 10^7 \text{ g}^{-1} \text{ soil})$ (d.f. = 4, P < 0.001; Fig. 2a). In contrast, at high altitude, archaeal *amo*A abundance $(4.63 \times 10^3 - 2.72 \times 10^4 \text{ g}^{-1} \text{ soil})$ was less than that for bacteria $(5.41 \times 10^4 - 3.1 \times 10^6 \text{ g}^{-1} \text{ soil})$ (d.f. = 6, P < 0.001; Fig. 2a). As a consequence, the mean \log_{10} ratio of archaeal: bacterial amoA decreased from 1.45 in low-altitude soils to -1.34 in high-altitude soils (Fig. 2b), with significantly negative and positive correlations, respectively, with altitude and archaeal amoA abundance (Table 2). There is a

Table 2. Correlation analyses among the abundance of AOA/AOB, altitude, pH, organic matter and total nitrogen content

| | Kendall's correlation coefficient | | | | | | |
|-------------------------|-----------------------------------|---------------------|----------------------------|----------|---------|-------------------|--|
| | Abundance of AOA | Abundance of AOB | Log ratio of AOA to AOB | Altitude | рН | Organic matter | |
| Abundance of AOB | 0.667** | | | | | | |
| Log ratio of AOA to AOB | 0.606** | 0.273 | | | | | |
| Altitude | - 0.543** | - 0.419 | - 0.605** | | | | |
| рН | - 0.377 | - 0.377 | - 0.233 | 0.312 | | | |
| Organic matter | 0.606** | 0.515** | 0.394 | - 0.326 | 0.664** | | |
| Total nitrogen content | 0.788** | 0.697** | 0.394 | - 0.388 | 0.521* | 0.576** | |

*Correlation is significant at the 0.05 level (two-tailed).

**Correlation is significant at the 0.01 level (two-tailed).



Fig. 2. Abundance of the bacterial and archaeal *amo*A along the altitude gradient of Mount Everest on the Tibetan Plateau. (a) The abundance of archaeal and bacterial *amo*A; (b) log₁₀ ratio of archaeal al: bacterial *amo*A abundance.

significantly negative correlation between the abundance of AOA and altitude, and a nonsignificant negative correlation between the abundance of AOB and altitude (Table 2). The abundance of AOA was positively correlated with the abundance of AOB, and significantly positive correlations were also observed between both AOA and AOB abundance and organic matter and total nitrogen (Table 2).

Community composition of AOA and AOB

Phylogenetic trees of the bacterial and archaeal *amo*A gene sequences and related NCBI sequences are shown in Figs 3 and 4. All AOB sequences fell within *Nitrosospira amo*A clusters 3a.1, 3a.2, ME, cluster 2-related and *Nitrosomonas amo*A clusters 6 and 7 (Avrahami *et al.*, 2002; Avrhami & Conrad, 2003). Sequences in cluster 3a fell within two subclusters 3a.1 and 3a.2 with bootstrap value at 90, and two distinct clades were designated as cluster ME and cluster 2-related. Cluster ME was related to, but outside of, cluster 1 and 4 groups with a bootstrap value of 90 and cluster 2-related sequences fell outside of cluster 2.

The proportion of sequence types representing various clusters in each clone library was calculated based on RFLP analysis of clones. In lower altitude soils (M1-M5), AOB communities were dominated by representatives of clusters 3a.1 and 3a.2, constituting 91.8-100% in each clone library, and also contained representatives of cluster 6 (Fig. 5a). The proportions of cluster 3a.1 and 3a.2 decreased in higher altitude soils (M6-M12) and only cluster 3a.1 was detected in two of seven high-altitude soils, in a low proportion (2.5-18.8%), while the cluster ME, cluster 2-related and cluster 6 sequences appeared in high proportions in these soils (Fig. 5a). Sequences falling within cluster ME and cluster 2-related groups were exclusively from the higher altitude soils (M6-M12) and accounted for 30.3-81.3% of the total clones in each clone library. Cluster ME sequences had < 94% similarity to the amoA-like sequences in GenBank database, with the exception of several clones, including AOB M10 20, AOB M11 29, AOB M6 18 and AOB_M12_10, which shared 98% similarity to the unclassified sequence AY667607 from plant rhizosphere soil (Hawkes et al., 2005). Most sequences affiliated with the cluster 2-related group shared < 94% similarity to sequences in GenBank database. The proportion of cluster 6 sequences was greater in higher altitude soils. Cluster 7 sequences were detected only in some higher altitude soils. In summary, therefore, the data indicate substantial differences in bacterial ammonia oxidizer community composition between the lower and the higher altitude soils.

All AOA sequences, with the exception of AOA M11 1, fell within soil/sediment lineage that was dominated by sequences from soil and sediment environments and was grouped into three distinct clades (Fig. 4). Clades 1 and 2 formed a monophyletic lineage with modest bootstrap support, and clade 3 was related to, but outside this monophyletic clade, which showed a topology similar to that observed by Hansel et al. (2008) for sequences from a continuous soil profile (Fig. 4). These were designated as soil and sediment clusters 1, 2 and 3 (Fig. 4). Soil and sediment clusters 1 and 2 represented the majority of all AOA clones, were present in all soils at varying proportions and corresponded to Saturated C Horizon and partial Mixed B and C Horizon clusters of Hansel et al. (2008), respectively (Fig. 5b). In lower altitude soils, more soil and sediment cluster 1 sequences (70-91.2%) and less cluster 2 sequences (5.8-22.5%) were obtained, while more cluster 2 sequences appeared with proportions of 31.8-92.3% in higher altitude soils (Fig. 5b). These sequences are most closely related to amoA sequences retrieved from various soils (Leininger et al., 2006; Chen et al., 2008; Hansel et al., 2008; Shen et al., 2008), alpine soils (G.W. Nicol & J.I. Prosser, unpublished data), sediments (Francis et al., 2005; Jiang et al., 2009), soil fosmid 54d9 (Treusch et al., 2005) and newly cultivated strain Candidatus Nitrososphaera gargensis



Fig. 3. Neighbor-joining phylogenetic tree of bacterial *amoA* sequences (465-bp fragment) retrieved from Mount Everest, on the Tibetan Plateau. Clones from this study are shown in bold with the name AOB, followed by soil code (M1–M12), and then by clone code. The accession number follows each clone. Bootstrap values (> 50) are indicated at branch points.



Fig. 4. Neighbor-joining phylogenetic tree of archaeal *amoA* sequences (594-bp fragment) retrieved from Mount Everest, on the Tibetan Plateau. Clones from this study are shown in bold with the name AOA, followed by soil code (M1–M12), and then by clone code. The accession number follows each clone. Bootstrap values (> 50) are indicated at branch points.

(Hatzenpichler *et al.*, 2008). Sequences falling within soil and sediment cluster 3 were present in four of five lower altitude soils with proportions of 1.8–22%.

The sequence AOA_M11_1, representing 24% of soil clones at an altitude of 6450 m a.s.l., fell within sediment and soil V cluster in marine lineage, which was previously characterized as being dominated by marine-originating sequences, but was grouped into two distinct clades (sediment and soil V cluster and marine cluster). The sediment and soil V cluster, defined by Nicol *et al.* (2008), was distinct from the marine cluster and mainly from soil and sediment environments (Fig. 4).

Discussion

Archaeal amoA abundance was markedly higher than bacterial amoA abundance in lower altitude soils (M1-M5). The result was consistent with many studies showing greater abundance of archaeal than bacterial amoA genes in most ecosystems (Leininger et al., 2006; Wuchter et al., 2006; He et al., 2007; Lam et al., 2007; Mincer et al., 2007; Nicol et al., 2008; Shen et al., 2008; Di et al., 2009). Intriguingly, archaeal amoA abundance decreased by up to four orders of magnitude between lower altitude and higher altitude soils, while bacterial amoA abundance decreased only 10-fold. As a result, the relative abundance of AOA and AOB shifted drastically, with AOB outnumbering AOA by 52 times on average in higher altitude soils (M6-M12). Variation in this ratio has also been described for coastal aquifer sediments, with greater abundance of AOA over AOB at low, but not high salinity (Mosier & Francis, 2008; Santoro et al., 2008), and in oxic lake water, but not in anoxic sediment (Jiang et al., 2009). Caffrey et al. (2007) also found that AOB were more abundant than AOA in several estuarine sites. The log₁₀ ratio of AOA to AOB correlated negatively with altitude and positively with AOA abundance, and AOA abundance was negatively related to altitude, indicating that AOA may be more sensitive than AOB in response to elevated soil conditions. Both AOA abundance and AOB abundance were positively correlated to organic matter and total nitrogen, suggesting that these may also determine the distribution of AOA and AOB in alpine soils, by providing the nutrient and substrate for microbial mineralization. It was also found that warming at high-elevation sites was more pronounced than at low elevations (Giorgi et al., 1997), and thus might exert different impacts on soil microbial populations for the low and high altitudes.

The abundance of AOA and AOB decreased considerably in soil at an altitude of 5700 m a.s.l. and then maintained a lower level in higher altitude soils. A corresponding shift was observed in AOB community composition between the lower altitude and the higher altitude soils. In lower altitude soils, the AOB community was dominated by *Nitrosospira*



Fig. 5. Proportions of bacterial *amoA* (a) and archaeal *amoA* (b) gene clone types representing various clusters in soils from Mount Everest, on the Tibetan Plateau.

cluster 3a, whereas two newly designated *Nitrosospira* cluster ME and cluster 2-related sequences and *Nitrosomonas* cluster 6 were prevalent in the higher altitude soils. The permanent snow line in the northern slope of Mount Everest is at 5800–6000 m a.s.l. Soils above or near this altitude are permafrost, representing a harsher habitat with altitude-dependent physiochemical gradients including frozen temperature, stronger radiation, lower oxygen concentration and lower nutrients, which may determine the abundance and community structure of AOA/AOB in elevated soils.

Among these altitude-dependent factors, temperature has been clearly demonstrated to have a huge influence on the community structure of AOB and caused apparent population shifts within the different *Nitrosospira* clusters (Avrahami *et al.*, 2003). Decreasing density of cluster 3a bands in denaturing gradient gel electrophoresis in response to a decrease in temperature, and a cluster 3a-dominated community at a high temperature were observed in previous studies (Avrhami & Conrad, 2003; Avrahami *et al.*, 2003). This is consistent with the predominance of cluster 3a in lower altitude soils and lack of detection of this group in most high-altitude soils, suggesting a response to temperature. In contrast, *Nitrosospira* cluster ME and cluster 2related sequences were most related to clusters 1, 2 and 4 and were present exclusively in higher altitude soils at high proportions. Clusters 1, 2 and 4 have so far only been found in cold temperate soils (Avrahami & Conrad, 2005). In soil microcosms, cluster 1 was found to be prevalent at a low temperature (4–10 °C), but disappeared completely at a high temperature (Avrahami *et al.*, 2003). High proportions of *Nitrosomonas* cluster 6 were also observed in high-altitude soils. Studies based on the pure culture *Nitrosomonas* cluster solw at temperatures as low as -5 °C (Jones *et al.*, 1988). Considering all these aspects, sequences belonging to cluster ME, cluster 2-related and *Nitrosomonas* cluster 6 may be representatives of a cold-tolerant AOB population and may be responsible for the observed higher abundance of AOB over AOA in these soils.

All AOA sequences fell within the soil and sediment clusters, suggesting that AOA communities in these alpine soils are distinct from those in aquatic environments. The proportions of the major cluster of AOA within clone libraries changed between lower and higher altitude soils, but no distinct horizon/site/treatment-specific clusters were observed as in the continuous soil profile (Hansel *et al.*, 2008), in estuary (Mosier & Francis, 2008) and agricultural soil (Nicol *et al.*, 2008; Di *et al.*, 2009), where pH is acidic or near neutral. Pronounced AOA community changes in

response to long-term fertilization treatments were observed in acidic soil (He *et al.*, 2007), but not in alkaline soils with pH ranging from 8.3 to 8.7 (Shen *et al.*, 2008). Considering the high soil pH values (8.6–9.1) in this study, together with the above studies, no remarkable shift in the AOA community between lower altitude and high-altitude soils may be attributed to the high pH in the Tibetan Plateau.

In summary, distinct changes in the abundance and community composition of AOB were observed at altitudes near or above the permanent snow line, suggesting that altitude-dependent factors, in particular temperature, have a profound influence on the abundance and community of AOB. AOA abundance decreased significantly and was negatively correlated with altitude in this study. Bacterial amoA abundance was greater than archaeal amoA abundance in higher altitude soils, contrary to observations in many systems. This distinct change in the relative abundance of bacterial and archaeal ammonia oxidizers, and in the bacterial ammonia oxidizer communities in response to elevated altitude also suggests that these groups are sufficiently active to survive in these alpine areas. More recent studies have demonstrated that microbial metabolism proceeds at soil temperatures as low as -20 °C (Christner, 2002), and that CO₂ production, CH₄ uptake and N cycling occur in subnival soil (Sommerfeld et al., 1993; Clein & Schimel, 1995; Brooks et al., 1996; Lipson et al., 1999), which supports this hypothesis. The ecological significance of the shifts in the relative abundance of AOA and AOB in these soils requires further study.

Acknowledgements

This work was supported by the Natural Science Foundation of China (40601049, 30811130224 and 40871129).

Authors' contribution

L.-M.Z. and M.W. contributed equally to this work.

References

- Agogue H, Brink M, Dinasquet J & Herndl GJ (2008) Major gradients in putatively nitrifying and non-nitrifying Archaea in the deep North Atlantic. *Nature* **456**: 788–791.
- Avrhami S & Conrad R (2003) Patterns of community change among ammonia oxidizers in meadow soils upon long-term incubation at different temperatures. *Appl Environ Microb* **69**: 6152–6164.
- Avrahami S & Conrad R (2005) Cold-temperate climate: a factor for selection of ammonia oxidizers in upland soil? *Can J Microbiol* **51**: 709–714.
- Avrahami S, Conrad R & Braker G (2002) Effect of soil ammonium concentration on N_2O release and on the

community structure of ammonia oxidizers and denitrifiers. *Appl Environ Microb* **68**: 5685–5692.

- Avrahami S, Liesack W & Conrad R (2003) Effects of temperature and fertilizer on activity and community structure of soil ammonia oxidizers. *Environ Microbiol* 5: 691–705.
- Bremner JM (1996) Nitrogen total. Methods of Soil Analysis: Part 3 – Chemical Methods, Vol. 37 (Sparks DL et al., eds), pp. 1085–1121. Soil Science Society of America, American Society of Agronomy, Madison, WI.
- Brooks PD, Williams MW & Schmidt SK (1996) Microbial activity under alpine snow packs, Niwot Ridge, Colorado. *Biogeochemistry* **32**: 93–113.
- Caffrey JM, Bano N, Kalanetra K & Hollibaugh JT (2007) Ammonia oxidation and ammonia-oxidizing bacteria and archaea from estuaries with differing histories of hypoxia. *ISME J* 1: 660–662.
- Chen XP, Zhu YG, Xia Y, Shen JP & He JZ (2008) Ammoniaoxidizing archaea: important players in paddy rhizosphere soil? *Environ Microbiol* **10**: 1978–1987.
- Christner BC (2002) Incorporation of DNA and protein precursors into macromolecules by bacteria at – 15 degrees C. *Appl Environ Microb* **68**: 6435–6438.
- Clein JS & Schimel JP (1995) Microbial activity of tundra and taiga soils at subzero temperatures. *Soil Biol Biochem* **27**: 1231–1234.
- Di HJ, Cameron KC, Shen JP, Winefield CS, O'Callaghan M, Bowatte S & He JZ (2009) Nitrification driven by bacteria and not archaea in nitrogen-rich grassland soils. *Nature Geoscience* **2**: 621–624.
- Francis CA, Roberts KJ, Beman JM, Santoro AE & Oakley BB (2005) Ubiquity and diversity of ammonia-oxidizing archaea in water columns and sediments of the ocean. *P Natl Acad Sci USA* **102**: 14683–14688.
- Giorgi F, Hurrell JW, Marinucci MR & Beniston M (1997) Elevation dependency of the surface climate change signal: a model study. *J Climate* **10**: 288–296.
- Hansel CM, Fendorf S, Jardine PM & Francis CA (2008) Changes in bacterial and archaeal community structure and functional diversity along a geochemically variable soil profile. *Appl Environ Microb* 74: 1620–1633.
- Hatzenpichler R, Lebecleva EV, Spieck E, Stoecker K, Richter A, Daims H & Wagner M (2008) A moderately thermophilic ammonia-oxidizing crenarchaeote from a hot spring. *P Natl Acad Sci USA* **105**: 2134–2139.
- Hawkes CV, Wren IF, Herman DJ & Firestone MK (2005) Plant invasion alters nitrogen cycling by modifying the soil nitrifying community. *Ecol Lett* **8**: 976–985.
- He JZ, Shen JP, Zhang LM, Zhu YG, Zheng YM, Xu MG & Di H (2007) Quantitative analyses of the abundance and composition of ammonia-oxidizing bacteria and ammoniaoxidizing archaea of a Chinese upland red soil under longterm fertilization practices. *Environ Microbiol* **9**: 2364–2374.
- Horz HP, Barbrook A, Field CB & Bohannan BJM (2004) Ammonia-oxidizing bacteria respond to multifactorial global change. *P Natl Acad Sci USA* **101**: 15136–15141.

Jiang H, Dong H, Yu B, Lv G, Deng S, Berzins N & Dai M (2009) Diversity and abundance of ammonia-oxidizing archaea and bacteria in Qinghai Lake, northwestern China. *Geomicrobiol J* **26**: 199–211.

Jones RD, Morita RY, Koops HP & Watson SW (1988) A new marine ammonium-oxidizing bacterium, *Nitrosomonas-cryotolerans* Sp-Nov. *Can J Microbiol* **34**: 1122–1128.

Könneke M, Bernhard AE, de la Torre JR, Walker CB, Waterbury JB & Stahl DA (2005) Isolation of an autotrophic ammoniaoxidizing marine archaeon. *Nature* **437**: 543–546.

Lam P, Jensen MM, Lavik G, McGinnis DF, Muller B, Schubert CJ, Amann R, Thamdrup B & Kuypers MMM (2007) Linking crenarchaeal and bacterial nitrification to anammox in the Black Sea. *P Natl Acad Sci USA* **104**: 7104–7109.

Leininger S, Urich T, Schloter M, Schwark L, Qi J, Nicol GW, Prosser JI, Schuster SC & Schleper C (2006) Archaea predominate among ammonia-oxidizing prokaryotes in soils. *Nature* **442**: 806–809.

Lipson DA, Schmidt SK & Monson RK (1999) Links between microbial population dynamics and nitrogen availability in an alpine ecosystem. *Ecology* **80**: 1623–1631.

Mincer TJ, Church MJ, Taylor LT, Preston C, Kar DM & DeLong EF (2007) Quantitative distribution of presumptive archaeal and bacterial nitrifiers in Monterey Bay and the North Pacific Subtropical Gyre. *Environ Microbiol* **9**: 1162–1175.

Mosier AC & Francis CA (2008) Relative abundance and diversity of ammonia-oxidizing archaea and bacteria in the San Francisco Bay estuary. *Environ Microbiol* **10**: 3002–3016.

Nicol GW, Leininger S, Schleper C & Prosser JI (2008) The influence of soil pH on the diversity, abundance and transcriptional activity of ammonia oxidizing archaea and bacteria. *Environ Microbiol* **10**: 2966–2978.

Park HD, Wells GF, Bae H, Criddle CS & Francis CA (2006) Occurrence of ammonia-oxidizing archaea in wastewater treatment plant bioreactors. *Appl Environ Microb* **72**: 5643–5647.

Prosser JI & Nicol GW (2008) Relative contributions of archaea and bacteria to aerobic ammonia oxidation in the environment. *Environ Microbiol* **10**: 2931–2941.

Reigstad LJ, Richter A, Daims H, Urich T, Schwark L & Schleper C (2008) Nitrification in terrestrial hot springs of Iceland and Kamchatka. *FEMS Microbiol Ecol* **64**: 167–174.

Rotthauwe JH, Witzel KP & Liesack W (1997) The ammonia monooxygenase structural gene amoA as a functional marker:

molecular fine-scale analysis of natural ammonia-oxidizing populations. *Appl Environ Microb* **63**: 4704–4712.

- Santoro AE, Francis CA, de Sieyes NR & Boehm AB (2008) Shifts in the relative abundance of ammonia-oxidizing bacteria and archaea across physicochemical gradients in a subterranean estuary. *Environ Microbiol* **10**: 1068–1079.
- Shen JP, Zhang LM, Zhu YG, Zhang JB & He JZ (2008) Abundance and composition of ammonia-oxidizing bacteria and ammonia-oxidizing archaea communities of an alkaline sandy loam. *Environ Microbiol* **10**: 1601–1611.
- Sommerfeld RA, Mosier AR & Musselman RC (1993) CO₂, CH₄ and N₂O flux through a Wyoming snowpack and implications for global budgets. *Nature* **361**: 140–142.

Tamura K, Dudley J, Nei M & Kumar S (2007) MEGA4: molecular evolutionary genetics analysis (MEGA) software version 4.0. *Mol Biol Evol* **24**: 1596–1599.

- Tourna M, Freitag TE, Nicol GW & Prosser JI (2008) Growth, activity and temperature responses of ammonia-oxidizing archaea and bacteria in soil microcosms. *Environ Microbiol* **10**: 1357–1364.
- Treusch AH, Leininger S, Kletzin A, Schuster SC, Klenk HP & Schleper C (2005) Novel genes for nitrite reductase and Amorelated proteins indicate a role of uncultivated mesophilic crenarchaeota in nitrogen cycling. *Environ Microbiol* 7: 1985–1995.

Venter JC, Remington K, Heidelberg JF *et al.* (2004) Environmental genome shotgun sequencing of the Sargasso Sea. *Science* **304**: 66–74.

Weidler GW, Dornmayr-Pfaffenhuemer M, Gerbl FW, Heinen W & Stan-Lotter H (2007) Communities of Archaea and Bacteria in a subsurface radioactive thermal spring in the Austrian Central Alps, and evidence of ammoniaoxidizing Crenarchaeota. *Appl Environ Microb* 73: 259–270.

- Wu SH, Yin YH, Zheng D & Yang QY (2007) Climatic trends over the Tibetan Plateau during 1971–2000. *J Geographical Sciences* **17**: 141–151.
- Wuchter C, Abbas B, Coolen MJL *et al.* (2006) Archaeal nitrification in the ocean. *P Natl Acad Sci USA* **103**: 12317–12322.
- Xu ZH, Chen CR, He JZ & Liu JX (2009) Trends and challenges in soil research 2009: linking global climate change to local longterm forest productivity. *J Soil Sediment* **9**: 83–88.